Spatio-temporal Variations and Potential Health Risks of Heavy Metals in Water from River Manafwa, Uganda

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Abstract: The epicenter of flash flood inundations and landslides in Uganda has been areas around Mt. Elgon. By implication, it has led to the loss of lives and food and water insecurity. This study assessed the seasonal variations in physiochemical parameters and heavy metals (HMs) water content from River Manafwa (R. Manafwa), the major water source used around Mt. Elgon. Potential insidious human health risks associated with consumption and dermal contact with water from the river were assessed using target hazard quotient and incremental lifetime cancer risk methods. Results of atomic absorption spectrometry analysis showed that the concentrations of the HMs in the wet and dry seasons ranged from below the detection limit to 1.407 ± 0.001 mg/L, which were below WHO limits. Health risk assessments indicated that there are discernable non-carcinogenic health risks from the ingestion of water from R. Manafwa, as the total target hazard quotients were above 1 for some of the samples. Cancer risk values indicated that there are no potential cancer risks from ingestion of water from the river. This study recommends that regulatory authorities should intervene to mitigate the pollution of R. Manafwa by strengthening restrictions on sand mining and dumping of wastes into the river.

Keywords: water quality; Manafwa watershed; Bududa; carcinogenic risk; target hazard quotient; trace metals.

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1. Introduction

Rapid economic growth, industrialization, and urbanization with inadequate regulatory structures have infringed on the pristine environment, raising many sustainability challenges [1]. A case in point is developing countries such as China and India, where pollution levels have surpassed regulatory limits and have been the subject of obsessive research [2, 3]. Of immediate concern has been the pollution of water resources with contaminants that make them unsafe for drinking, thereby impeding the realization of some Sustainable Development Goals such as SDG 6 and SDG 14 [4]. Over the years, the scope of water contaminants has widened to include heavy metals (HMs), endocrine-disrupting chemicals, flame retardants, polycyclic aromatic hydrocarbons, current use pesticides, preservatives, personal care products [5-7], and other contaminants of emerging concern such as macro-, micro- and nanoplastics, algal toxins, and active pharmaceutical ingredients [7]. Within this frame of reference, HMs remain the most

ubiquitous contaminants in water resources due to their diverse occurrence and inclusion in appliances, consumer goods, industrial machinery and processes [8].

Heavy metals (HMs) are chemical elements with relatively high densities and are potentially toxic at concentrations above their established threshold limits [9]. They form part of the earth's crust in negligible concentrations but can continuously get enriched due to indiscriminate human activities (that alters their geochemical cycles and biochemical balance) and natural processes such as volcanic activity, metal corrosion, and metal evaporation from soil, water, and sediments [8, 10]. Examples of HMs include typical metals and metalloids such as vanadium, tin, strontium, mercury, arsenic, lead, zinc, nickel, cadmium, chromium, cobalt, copper, iron, molybdenum, and titanium [9, 11]. Some of the HMs are important co-enzymes in reactions that drive living cells but are bioaccumulative and toxic to living organisms at elevated concentrations [12, 13]. For terrestrial and aquatic organisms, exposure to HMs occurs through direct ingestion in water, food, medicine, direct deposition from the atmosphere, dermal adsorption, or inhalation from occupational sources [8, 14-16]. Ingestion of HMs in drinking water is the most common route of exposure, morbidities, and mortalities from HMs [17]. Despite this, routine monitoring of HMs contamination and remediation in water resources of developing countries are limited.

There are various water resources in the East African community, a region of seven developing countries in Eastern Africa. Nevertheless, there are still high incidences of water scarcity in the region [18, 19]. For example, Uganda, the study's focus area, has over 21 million (51%) of its population without access to safe drinking water. This is in part due to the contamination of the available water resources by various anthropogenic activities. Around Mt. Elgon (Manafwa watershed) in Eastern Uganda, there have been several incidences of torrential rains, perennial (flash) floods, and landslides, which have led to an intolerable death toll, food, and water insecurity [20, 21]. The floods have spotlighted the country's climate change crisis, emphasizing the need for more research in this watershed [22].

To date, no study has examined the HMs content as well as health risks that could emanate from ingestion and dermal contact with contaminated water from River Manafwa (also known as Manafwa River; R. Manafwa henceforth) especially following torrential rains and the attendant floods and landslides. Therefore, this contribution assessed the seasonal variations in some selected physiochemical parameters and assessed the health risks associated with ingestion and dermal contact with HMs in water from R. Manafwa, the major water source used across Bududa, Manafwa, Butaleja, and Mbale districts of Eastern Uganda.

2. Materials and Methods

2.1. Description of the study area.

The study was undertaken on water samples from R. Manafwa (0.9420°N 33.920°E), the longest river that flows through the Mbale and Butaleja districts of Uganda. The river is fed by various tributaries (such as Sala, Liisi, Wukha, Tsutsu, Pasa, Kufu, Nambale, and Makhuba) and small streams from the transboundary Mt. Elgon (an extinct volcanic agglomerate of Miocene age that rises to 4321 m above sea level)[23]. River Manafwa is about 14.63 kilometers long and is the main water source for the nearby communities. Still, economic activities (such as sand mining, agriculture, and welding works) and frequent floods have anecdotally been implicated in the deterioration of the water quality of R. Manafwa that forms part of the Manafwa River Basin traversing Bududa, Manafwa, Butaleja, and Mbale districts

of Eastern Uganda (Figure 1) [24]. Rivers and streams from Mt. Elgon densely dissect the Manafwa River Basin, explaining the rugged topography characterizing the steep slopes in this area. Hydrologically, R. Manafwa joins the Mpologoma River and drains into Lake Kyoga, a mesotrophic lake connected to Lake Victoria, and forms part of the River Nile that flows up to Egypt and empties into the Mediterranean Sea [25].



Figure 1. Map showing the location of River Manafwa in Eastern Uganda.

River Manafwa is the major river in the Elgon sub-region. It is the major water source for Manafwa waterworks that supplies water for domestic use to Mbale municipality, Butaleja town, and Kamonkoli town in Budaka district. Its water is also used by the local community for agriculture along the river banks, fish farming and rice growing in the Doho irrigation scheme, Butaleja district. This study considered R. Manafwa stretching from Bulucheke in Bududa district up to slightly beyond Manafwa Water Works, Mbale district (Figure 1). This covered a total distance of about 58 km. The river was divided into two sections, namely, the upper course and the lower course. The upper course stretched from Bulucheke to Manafwa town. This part of the river was further subdivided into two segments, A and B (Table 1). Segment A was from Bulucheke to Bududa, and this was an area where landslides normally occur. Segment B was from Bududa Town up to Manafwa Town, an area characterized by great human settlements and activity in the river catchment area. The lower course of the river was from Manafwa town up to slightly beyond Manafwa Water Works (Figure 2); with the greatest human activities including stone quarrying, sand mining, agriculture, automobile washing, and vehicular movement and sewage treatment. Besides, this part of the river also receives road runoff from Mbale Municipality and Manafwa Town, where many metals work activities and leachates from solid waste heaps occur.

Table 1. Summary of sampling sites and the potential sources of heavy metal contaminants along R. Manafwa.Sampling SiteSample codeActivities/establishments

Samping Site	Sample code	Activities/establishments
Bulucheke area	A1, A2	Landslide affected site, residential area, school science laboratories
Bukigai market area	B1	Market activities, garages, residential cultivation
Bududa Town	B2	Metal welding and fabrication, Bududa Hospital, car washing bay
Manafwa Town downstream	M1	Metal welding, garages, horticulture cultivation, slum and residential

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Figure 2. Overview of R. Manafwa (a) river water choked with high levels of sand and silt; (b) staff at Manafwa Water Works trying to clean water clarifiers to avoid them from silting up, (c) one of the sand mining site operations that is speeding up river bank erosion and sedimentation, and (d) section of the river where flooding swept away a bridge.

2.2. Sample collection.

The water samples were collected at a depth of 10-15 cm below the water surface, using clean 500 mL plastic bottles. Sampling was done at 8.00 am, 12.00 pm, and 4.00 pm (East African Standard Time) to cater for variations that could have occurred in the river water due to temperature changes and the different human activities occurring during the day. Sampling was done thrice each season (September, October, and November 2018 for the wet season, and December 2018, January 2019, and February 2019 for the dry season).

All three water samples collected from each sampling site were mixed to obtain a composite water sample. One liter of the composite water sample was measured and transferred into a cleaned plastic bottle, sealed and labeled to easily identify the sampling site. Four composite water samples were taken from each of the two river sections during each sampling in the major seasons (dry and wet seasons) to account for seasonal variations. In the river's upper course, two composite water samples were taken from each segments A and B and labeled A1, A2, and B1, B2, respectively. In the river's lower course, two water samples were taken from two selected sites within the area just before the Manafwa water works and were labeled M1 and M2. Two other water samples were taken from sites just after the bridge and beyond Manafwa water, labeled M3 and M4. The eight sampling points were chosen to capture the stretch where the landslides normally occur and the major anthropogenic activities carried https://nanobioletters.com/

out along the river banks (Table 1). Water samples were then transported to the laboratory in ice boxes within 24 hours from the time of collection.

2.3. Sample preparation and analyses.

During the sampling of water, non-conservable parameters (electrical conductivity and pH) were measured on-site using calibrated Jenway pH/mV/Temperature and Conductivity meters (Jenway Gransmore Green, England). The other parameter (total hardness) was determined by EDTA complexometric titration method [26].

Each water sample was filtered through Whatmann No. 42 filter paper to remove suspended solids. They were then acidified with 1 ml of concentrated nitric acid to preserve ions in solution in preparation for elemental analyses. The acidified water samples (250 mL) were evaporated into 25 mL volumes. The concentrates were transferred into 50 mL volumetric flasks and diluted to the mark with diluents (0.1% of lanthanum chloride and 1% of concentrated nitric acid in distilled water). These were then transferred into clean sample bottles and analyzed for the HMs using atomic absorption spectrophotometer (AAS, Perkin Elmer Analyst 100) to determine the concentration of the eight selected HMs: copper (Cu), nickel (Ni), manganese (Mn), zinc (Zn), lead (Pb), cadmium (Cd), chromium (Cr) and iron (Fe).

Working standards prepared from the dilution of 1000 ppm stock solution of the nitrate and chloride salts of the HMs were used to construct calibration curves. The concentration of the HMs in the sample digestates was determined from the calibration curves in mg/L. Quality control was performed by analyzing blanks and spiked samples according to the same procedure. Recoveries obtained ranged from 96% to 101%. Analytical precision (expressed as Relative Standard Deviation) varied between 3% and 4%. The method detection limits (LODs) were computed as Blank + $3 \times$ Standard Deviations for four samples analyzed in triplicate.

2.4. Assessment of human health risks.

Health risk assessments establish the link between the environment and human health that can be expressed quantitatively in terms of hazard degree. This study calculated the carcinogenic and non-carcinogenic health risks separately for adults as the general population and children as a sensitive group.

2.4.1. Non-carcinogenic health risks

The average daily doses were estimated to discern human exposure through direct ingestion (ADD_{ingestion}; mg/kg/day) and dermal contact (ADD_{dermal contact}; mg/kg/day) with water (Equations 1 and 2)[15, 27, 28]. Dermal contact is expected to stem from sand mining in this river, exposing the mining communities to HMs in the contaminated water.

$$ADD_{ingestion} = \frac{C \times W_{ir} \times E_f \times E_d}{W_{ab} \times T_{aet}}$$
(1)

$$ADD_{dermal\ contact} = \frac{C \times SAF \times DAF \times AF \times E_f \times E_d}{W_{ab} \times T_{aet}} \times 10^{-6}$$
(2)

From which C is the heavy metal concentration (mg/L), W_{ir} is the water ingestion rate = 1.8 L/day and 21.0 L/day for children and adults, E_f = exposure frequency (365 days/year), E_d = exposure duration, the average lifetime (58.65 years for an adult Ugandan)[15, 29], W_{ab} https://nanobioletters.com/

= average body weight (considered to be 15 kg for children and 60 kg for adults), T_{aet} is the average exposure time for non-carcinogens = $E_f \times E_d$ [30], SAF is the exposed surface area = 2,800 cm² for children and 24,350 cm² for adults [28], DAF is the dermal absorption factor = 0.01 for carcinogenic HMs and 0.001 for non-carcinogenic HMs [31], AF is the skin adherence factor in mg/cm²/day = 0.2 and 0.7 for children and adults [32].

Similarly, the target hazard quotient (THQ) was calculated for both direct ingestion and dermal contact with water (Equations 3 and 4). Practically, THQ ≤ 1 is indicative that the exposure is unlikely to elicit adverse health effects on an individual. Otherwise, THQ > 1 attests to the potential of non-carcinogenic effects being experienced [32]. Because such effects are augmentative in the context of contaminants like HMs, the cumulative risk or total THQ was computed as the arithmetic sum of the THQ of the HMs. As per US EPA [33], the health risk calculations assume that the ingested and adsorbed doses are equal to the dose absorbed into the body.

$$THQ = \frac{ADD_{ingestion}}{R_f D_{oral}}$$
(3)

$$THQ = \frac{ADD_{dermal contact}}{R_f D_{dermal}}$$
(4)

where $R_f D_{oral}$ is the oral reference dose. Its values are 4.0×10^{-2} , 3.0×10^{-4} , 3.0×10^{-2} , 3.0×10^{-2} , 4.0×10^{-3} , 1.0×10^{-3} , 1.5×10^{0} , and 7×10^{-1} mg/kg/day for Cu, Ni, Mn, Zn, Pb, Cd, Cr, and Fe, respectively. The $R_f D_{dermal}$ is the dermal reference dose, with values of 1.0×10^{-2} , 5.40×10^{-3} , 9.6×10^{-1} , 6.0×10^{-4} , 5.25×10^{-4} , 6.0×10^{-5} , 6.0×10^{-5} , and 1.4×10^{2} mg/kg/day for Cu, Ni, Mn, Zn, Pb, Cd, Cr and Fe, respectively [33]. A reference dose is defined as the highest amount of a heavy metal ingested through a given pathway which may not result in risk of developing deleterious health effects in an individual in their lifetime [34]. Thus, if the average daily dose (ADD_{ingestion} or ADD_{dermal contact}) is lower than the respective reference dose, the THQ < 1 and adverse health effects are unlikely. Otherwise, an average daily dose greater than the reference dose is indicative that THQ > 1 and adverse health effects may occur.

2.4.2. Cancer risk assessment.

The carcinogenic health risk (CR) estimated as the incremental lifetime cancer risk for the carcinogenic HMs (Pb, Cd, Cr, and As) was calculated as the product of ADD_{ingestion} and the ingestion cancer slope factor (CSF) using Equation 5. The total cancer risk (TCR) was calculated using Equation 6, used in previous studies [31, 35].

$$CR = ADD_{ingestion} \times CSF$$
 (5)

$$TCR = \sum_{i=1}^{n} CR \tag{6}$$

The CSF for Pb, Cr, and Cd are 8.5×10^{-6} , 5.0×10^{-4} , and 3.8×10^{-4} mg/kg/day, respectively. The CSF is defined as the risk generated by an average lifetime amount of one mg/kg/day of carcinogen chemical and is contaminant specific. The US EPA permissible limits lie between 10^{-6} and $< 10^{-4}$ for a single carcinogenic element and multi-element carcinogens [31].

2.5. Statistical analysis.

Quantitative data from triplicate analyses were entered into Excel, where they were averaged. Significant differences in the spatial variations in water quality among the sampling sites along the river were evaluated using the One Way Analysis of Variance (One Way ANOVA) with Tukey posthoc test. Pearson's bivariate correlation and Principal Component Analysis (PCA) were used to explore the inter-relationships between metal concentrations and the examined physicochemical parameters of R. Manafwa water. The analyses were executed at 95% confidence interval employing GraphPad Prism for Windows (v9.3.1, GraphPad Software, San Diego, CA, USA).

3. Results and Discussion

3.1. Variations in physicochemical parameters of R. Manafwa water.

Table 2 shows the results of pH, conductivity, and total hardness of water from the different sections of R. Manafwa during the wet and dry seasons. The parameters did not differ significantly among the study stations (p > 0.05) and seasons (p > 0.05). The pH (a measure of hydrogen ions) ranged from 7.46±0.05 in the wet season to 8.19±0.06 in the dry season and were within the acceptable limits of the World Health Organization (WHO) for drinking water [36]. These values were comparable to 7.96-8.22, 6.2-8.0, 6.6–7.5, 5.85-7.60, and 8.05-8.30 reported for water from Nyabugogo and Nyabarongo rivers, Rwanda [35], Mohokare River (Lesotho)[37], River Aturukuku [38], River Nyamugasani [39] and River Rwimi of Uganda [39, 40], but higher than 5.60-6.32 and 5.58-6.80 for River Nyamwamba and River Mubuku of Uganda [40]. Though the pH values obtained lie within acceptable limits, it has been indicated that even slightly high or low water pH is unpleasant. For example, high alkalinity confers a slippery feel to water, making it taste like baking soda, while at highly acidic pH, water possesses a bitter and/or metallic taste and induces corrosion [41]. At very low and high pH levels, the solubility of toxic HMs in water tends to increase and can cause serious human health effects in humans and aquatic organisms [42, 43].

Electrical conductivity (EC) was measured to establish the total dissolved ions in the water samples. It ranged from 88.7±0.24 µS/cm in the dry season to 122.2±0.91 µS/cm in the wet season, which was within the WHO guidelines for drinking water. The values obtained in this study are comparable to 80.44, 63.15 and 12–119, 43–103, and 99.91 µS/cm for water in Ugandan rivers: Lhubiriha, Mobuku, Rwimi, and Nyamwamba but lower than 460.51, 946.08, 118.57, 81–220 and 140.82 µS/cm in River Lubigi, River Nyamugasani, River Sio, River Rwimi and River Victoria Nile reported by Bwire et al. [44] and, Busulwa and Bailey [39]. Turinayo [45] reported EC of 108-1524 µS/cm for water from River Musamya in Uganda. In Nyabugogo and Nyabarongo rivers in Rwanda, EC of 74.3-102.0 were reported [35]. Another investigation in Mohokare River water (Lesotho) [37] reported EC of 2000-3800 µS/cm, which is far higher than obtained in this study. In Nigeria, Butu et al. [46] found the EC of River Rido to range from 79 to 146.3 μ S/cm [46], which is close to the ones obtained in this study. The EC of water estimates the total amount of solids dissolved in water (its total dissolved solids) and is directly proportional to the water's temperature. It is directly related to the concentration of ions in the water, and this is supported by the low levels of HMs reported in this study. High EC of water samples indicates the presence of a higher content of different salts and organic and inorganic materials such as alkalis, chlorides, sulfides, and carbonates. As a measure of water quality, significant changes in EC are indicators of discharges or some other source of pollution entering the river.

		wet season		Dry season				
Sampling site	рН	Conductivity (µS/cm)	Total hardness (mg/L)	рН	Conductivity (µS/cm)	Total hardness (mg/L)		
A ₁	7.95 ± 0.02	107.6±0.12	104.2±0.14	8.10±0.01	100.3±0.05	100.5±0.19		
A_2	8.14 ± 0.04	119.6±0.43	112.2±0.67	8.19 ± 0.06	112.0±0.11	108.7±0.40		
B1	7.84 ± 0.01	98.3±0.23	103.2±0.28	7.99±0.01	88.7±0.24	101.5 ± 0.01		
B ₂	7.82 ± 0.07	101.2±0.15	103.0±0.56	7.80±0.05	94.9±0.10	103.0±0.00		
M_1	7.74±0.01	101.9±0.45	115.9±0.76	7.79±0.01	89.5±0.22	106.1±0.52		
M_2	7.46 ± 0.05	106.3±0.76	103.4±0.27	7.50 ± 0.03	97.8±0.09	95.9±0.10		
M ₃	7.73±0.04	122.2±0.91	101.8±0.11	7.77±0.00	115.0±0.14	93.6±0.24		
M_4	8.09±0.01	119.2±0.17	89.4±0.13	8.00±0.02	101.6±0.10	80.8±0.11		
WHO guidelines [36]	6.5-8.5	250.0	200.0	6.5-8.5	250.0	200.0		

 Table 2. Physicochemical parameters of water samples from R. Manafwa, Eastern Uganda.

 Wet season

Values are means \pm standard deviations of analyses performed in triplicate.

The total hardness of the water samples ranged from 80.8 ± 0.11 mg/L in the dry season to 115.9 ± 0.76 mg/L in the wet season. Based on available criteria, hardness ranging from 80 to 100 mg/L is deemed to be optimal, while those more than 300 mg/L are potentially hard. Overall, water having a hardness of 75 mg/L or less is soft. Water exceeding 200 mg/L is considered poor, whereas those with hardness exceeding 500 mg/L are unacceptable for domestic purposes [47]. Highly hard water is chemically rich in dissolved minerals, especially calcium and magnesium ions which, for aesthetic reasons, may have an unpleasant taste [48]. Both low and high values of hardness may be harmful to the human body. Low levels of hardness may activate colon carcinogens or trigger rectal cancer and cardiovascular diseases [49, 50] because calcium and magnesium ions are capable of binding bile acids and fatty acids, thus affecting the creation of colon mucosa [51]. Higher hardness values may lead to the development of kidney stones and dermal diseases [52]. In addition to these health risks, hard water is a nuisance as it causes mineral buildup on fixtures (hence corrosion) and poor soap or detergent performance due to scum formation.

3.2. Spatio-temporal variations in heavy metal content of River Manafwa water.

3.2.1. Copper.

The seasonal fluctuations in Cu content of R. Manafwa water are shown in Figure 3. The concentrations of Cu ranged from $0.023\pm0.003 \text{ mg/L}$ to $0.06\pm0.01 \text{ mg/L}$ during the wet season and $0.0030\pm0.01 \text{ mg/L}$ to $0.0045\pm0.01 \text{ mg/L}$ in the dry season. The mean values were 0.0382 mg/L and 0.0034 mg/L for the wet and dry seasons. The concentration of Cu generally increased from the upper course of the river downstream in both seasons (p<0.05), but none of the values exceeded the WHO guidelines of 0.5 mg/L in drinking water [36]. The increasing concentration of Cu downstream could be due to runoff from Manafwa town and Mbale municipal motor garages, metal fabrication, road construction, and leachates from dumped domestic biodegradable wastes. The concentration of Cu was higher during the wet season, with $0.06 \pm 0.01 \text{ mg/L}$ at sampling point M4. This elevated concentration of Cu is probably due to road construction works that were going on during sampling at the Manafwa bridge and leaching from Mbale municipal sewage pipes. Such Cu leaching is accelerated by water characteristics such as high acidity and temperature, and low hardness [53].



Figure 3. The concentration of copper in water from River Manafwa in the wet and dry seasons. Values are means of analyses performed in triplicate (n = 16 composite samples).

In comparison to previous studies (Table 3), the concentrations of Cu reported in this study are quite lower. Though it is an essential trace metal, ingesting Cu at high concentrations leads to Cu poisoning, a condition characterized by hematemesis and gastrointestinal distress. Similar effects and neurological disorders such as Wilson's, Menkes', Alzheimer's, and Parkinson's diseases have been reported by individuals with dysregulation of the redox-active metal [53, 54].

Table 5. Con	ipanson or the		ivy metai wa	ter containina	tion (mg/L) non	i ix. manar	wa with pic	Tous studies	
River (Country)	Cu	Ni	Mn	Zn	Pb	Cd	Cr	Fe	References
Diver Menefive (Heende)	0.002.0.060	0.0015-	0.011-0.262	0.0015-0.029	0.002-0.010	0.001-	0.003-	0.196-	This study
Kiver Manarwa (Oganda)	0.003-0.000	0.0095				0.002	0.011	1.407	
Marimba River (Zimbabwe)	0.13-0.14	—	—	ND	0.213-0.544			5.6-6.9	[55]
Madanzhe, Dzindi and Mvudi rivers (South Africa)	0.002-0.003			0.0021- 0.0025	0.0105–0.0201	0.0016– 0.0093	_		[56]
River Sosiani (Kenya)	0.001-0.275			0.07-0.57	0.02-1.89	_	0.003- 0.050	0.011- 3.789	[57]
River Nyamwamba (Uganda)	1.90–61.0	0.67-12.0	23.1-100.0	ND ¹	0.27-0.40	—	—	185.0- 265.0	[58]
	0.740	_		0.076	0.047		_		
River Rwimi (Uganda)	0.010			0.010	0.067				[40]
River Mubuku (Uganda)	0.025			0.010	0.053				
Nyabarongo river (Rwanda)	BMDL-0.24		0.02-0.53	BMDL-0.09	0.05-0.75	BMDL- 0.106	BDL-0.06	0.63-1.61	[35, 59]
Bongna Piyer (China)	0.00189-	0.00745-	0.0431-	0.013 0.415	0.00049-	0.00012-	0.00156-		
Kongna Kivel (Clilla)	0.806	0.0601	2.041	0.013-0.413	0.00241	0.00064	0.00637		[60]
Bolong river (China)	0.00542-	0.0107-	0.0378-	0.0398-	0.00034-	0.00011-	0.00174-		[00]
Bolong itver (Clillia)	0.00737	0.0182	0.0519	0.567	0.00064	0.00054	0.00308		

Table 3 Comparison of the results of heavy metal water contamination (mg/L) from **R** Manafwa with previous studies

¹ BDML = Below method detection limit.

3.2.2. Nickel.

The concentration of Ni ranged from 0.0015 ± 0.001 mg/L to 0.0095 ± 0.005 mg/L during the wet season and 0.001 ± 0.010 mg/L to 0.005 ± 0.010 mg/L in the dry season (Figure 4). Like in the case of Cu, the concentration of Ni was higher (p>0.05) during the wet season, and samples were taken within Bududa and Manafwa towns (sampling sites B2, M1, and M2). The major sources could be the use of nickel-cadmium batteries, hydrogenated oils such as margarine, stainless steel, nickel-plated metallic items, and the combustion of petroleum fuels in Bududa and Manafwa towns. Sample A2, taken near Bukigai markets, had a higher concentration of Ni within the upper course of R. Manafwa. This could be due to market waste https://nanobioletters.com/

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and vehicular emissions due to heavy traffic flow during market days. In both the wet and dry seasons, the concentration of Ni in the water samples never surpassed the WHO maximum permissible limit of 0.05 mg/L for Ni in drinking water [36]. Previous studies in River Nyamwamba (Uganda) [58], Rongna, and Bolong rivers (China) [60] detected Ni at concentrations that are higher than was obtained in this study.



Figure 4. The concentration of nickel in water from R. Manafwa in the wet and dry seasons. Values are means of analyses performed in triplicate (n = 16 composite samples).

3.2.3. Manganese.

In this study, Mn concentrations varied between 0.011 ± 0.05 mg/L and 0.164 ± 0.01 mg/L during the wet season and 0.053 ± 0.007 mg/L and 0.262 ± 0.03 mg/L during the dry season (Figure 5). In the wet season, the concentration of Mn was generally lower (p<0.05) and showed an increase downstream, which could result from discharge from increasing human settlement at the banks of the river. Sample M4 taken downstream of Manafwa Water Works and Mbale-Tororo highway had elevated concentrations of Mn. Intense sand mining and construction repairs on Manafwa Bridge could be the cause. Evident metallic materials like spades and hoes used in sand mining and the combustion of petrol in motor cars could be the sources of Mn in the river water.

In the dry season, the concentration of Mn increased from the upper course of the river downstream and was higher than in the wet season (p<0.05). The source of Mn could be metallic materials like spades and hoes used in sand mining. Samples from sites M3 and M4 taken downstream of Manafwa Water Works with more intense sand mining showed the highest Mn concentrations. The concentration of Mn in the wet season was below the WHO maximum acceptable limit of 0.1 mg/L at all the study sites except site M4. In the dry season, it was above the maximum acceptable limits in all the sites except A1 and B2. These values were comparable to 0.02 mg/L to 0.53 mg/L reported in the Nyabugogo and Nyabarongo rivers, Rwanda [35, 59] but lower than 10.28 mg/L and 11.58 mg/L for water samples from Rwanzekuma and Ruganwa rivers of Rwanda [61].



Figure 5. The concentration of manganese in water from R. Manafwa in the wet and dry seasons. Values are means of analyses performed in triplicate (n = 16 composite samples).

3.2.4. Zinc.

On the other hand, Zn was found in levels of 0.008 ± 0.040 mg/L to 0.029 ± 0.001 mg/L in water samples collected during the wet season and 0.0015 ± 0.010 mg/L to 0.0045 ± 0.000 mg/L during the dry season (Figure 6). The highest concentration of 0.029 ± 0.001 mg/L was for samples from M4 in the wet season. The concentration of Zn in the wet season was higher than that in the dry season at all the sampling points (p<0.05). This could be due to the mobilization of the soluble forms of the metal from anthropogenic activities by runoff rainwater. In the wet season, water samples from A2 near the Bukigai market showed a higher concentration of Zn than those from sites A1, B1, B2, and M1. This could be attributed to the leachates from piles of biodegradable wastes generated from the market since Zn is one of the major micronutrients and vehicular emissions. None of the samples, however, contained Zn in levels surpassing the WHO limit of 3 mg/L in drinking water.



Figure 6. The concentration of zinc in R. Manafwa water samples collected during the wet and dry seasons. Values are means of analyses performed in triplicate (n = 16 composite samples).

3.2.5. Lead.

The seasonal fluctuations in Pb concentrations were more pronounced. The concentrations were from 0.002 ± 0.040 mg/L to 0.01 ± 0.005 mg/L in the wet season and 0.005 ± 0.02 to 0.0085 ± 0.010 mg/L in the dry season (Figure 7). The concentration of Pb decreased from the upper course of the river downstream to Bududa town in both the wet and dry seasons. This indicates the existence of a mild point source of Pb in the upper course of the river. Since there are hardly any settlements or economic activity evidenced at the uppermost course of R. Manafwa, the higher concentration of Pb at this section of the river could be due to natural sources such as landslides or hydro-geochemical reactions occurring rather than anthropogenic sources. The decrease in the concentration of Pb towards Bududa town is attributed to increased distance away from the source of pollution due to possible sedimentation and precipitation. From Bududa town (site B2) downstream up to slightly beyond Manafwa Bridge along Mbale-Tororo highway (site M4), the concentration of Pb increased. This could be due to vehicular emissions, metal welding and fabrication works, disposal of Pb acid battery contents, use of leaded gasoline [62, 63], and sewage leakages from Manafwa and Mbale towns.



Figure 7. The concentration of lead in R. Manafwa water samples. Values are means of analyses performed in triplicate (n = 16 composite samples).

The concentration of Pb reported in this study did not exceed the WHO maximum limit of 0.01 mg/L. Similar results have been documented for other rivers on the African continent. For example, Okonkwo and Mothiba [56] reported concentrations of 0.010 mg/L to 0.012 mg/L for Pb in water from Dzindi, Madanzhe, and Mvudi rivers, South Africa. Other studies, such as Amadi [57] in Sosiani River, Kenya, and Mvungi et al. [55] in Marimba River (Zimbabwe), reported higher Pb contents than those reported in this study (Table 3). Pb is toxic and a nonessential trace metal. Its chronic ingestion or occupational exposure has been associated with renal failure and liver degradation [64]. In infants, Pb retards interactive, survival, growth, development, and metabolic processes in addition to increasing mucus synthesis and triggering nervous system disorders [65].

3.2.6. Cadmium.

Cadmium was detected at relatively lower concentrations (0.001 ± 0.04 to 0.002 ± 0.00 mg/L) during the wet season but was not detected in the dry season (Figure 8). The highest

concentration of 0.002 mg/L was recorded in the water sample from M4. These results suggest that the Cd found in water samples for the wet season could have been mobilized by runoff water from metal fabrication works, disposed nickel-cadmium batteries, sewage, and other sources related to man's use of metals. The absence of Cd in the dry season could also be due to sedimentation/precipitation of the insoluble forms of Cd into sediments. In all the water samples, the cadmium concentration was below the WHO permissible limit of 0.03 mg/L.



Figure 8. The concentration of cadmium in water samples from R. Manafwa, Uganda, during the wet season. Values are means of analyses performed in triplicate (n = 8 composite samples). Dry season samples had no detectable Cd.

3.2.7. Chromium.

Unlike Cd, Cr was detected during the dry season at site M_4 at a concentration of 0.00093 ± 0.01 mg/L (Figure 9). For wet season samples, the concentration of Cr ranged from 0.003 ± 0.05 mg/L to 0.011 ± 0.01 mg/L for samples obtained near Manafwa water works (site M4). The relatively higher concentration of Cr at this sampling point could be due to releases from the heavy traffic flow along the Mbale-Tororo highway and leaching from the sewer pipes of Manafwa Water Works. By implication, the observed concentration of Cr during the wet season should be due to its mobilization from anthropogenic and geologic sources by rainwater.



Figure 9. The concentration of chromium in R. Manafwa water samples. Values are means of analyses performed in triplicate (n = 16 composite samples).

3.2.8. Iron.

In comparison to other HMs, Fe was found to be higher in the river water during both seasons, ranging from 0.967 ± 0.06 and 1.407 ± 0.001 mg/L in the dry season and between 0.196 ± 0.01 and 0.909 ± 0.03 mg/L in the wet season (Figure 10). The dry season had higher concentrations of Fe, possibly due to an increase in sand mining within the river when the water levels have reduced as a result of using metallic spades and hoes.



Figure 10. The concentration of iron in R. Manafwa water samples. Values are means of analyses performed in triplicate (n = 16 composite samples).

Inorganic Fe (principally Fe^{2+}) is naturally present in water resources at concentrations up to 50 mg/L [66]. For example, Amadi [57], Eliku, and Leta [67] found Fe (0.011 to 2.897 mg/L, and 1.11 to 4.12 mg/L) in water from River Sosiani (Kenya) and Awash River (Ethiopia) which are comparable to the values obtained in this study. However, the levels found in water in this study are lower than previously detected in some rivers. Nhapi et al. [61] reported elevated Fe levels (8.76 and 6.85 mg/L) in water from the Rusine and Marenge rivers of Rwanda, whereas Kihampa and Wenaty [68] detected Fe in Mara river water (Tanzania) at levels (12.6 to 15.51 mg/L) which are several folds higher than found in this study.

Taken together, the differences in HMs concentration reported in water from rivers across the globe and the present study could plausibly be due to disparities in the geological formation of the rivers, their physicochemical conditions, and the sources of contamination in their vicinities [60, 61].

3.3. Multivariate statistical analysis results.

To discern the pollution sources for the HMs, Pearson's correlation analysis and PCA were performed. In the wet season, there were positive correlations between Cu and Mn (p = 0.013), Cu and Zn (p = 0.012), Cr and Cu (p = 0.009), Pb and Mn (p = 0.015), Zn and Mn (p = 0.088), Cd and Mn (p = 0.040), and Cr and Mn (p = 0.000). The total hardness had strong and significant negative correlations with Cu (p = 0.011), Mn (p = 0.028), and Cr (p = 0.010) (Table 4). These observations agreed with PCA results (Figure 11). The strong correlation between the metal pairs suggests that they entered the river from anthropogenic sources or marked similarities in their chemistry [16, 69, 70]. For the negative correlations, it indicates the presence of the respective ions exerts a negative effect on hardness [71].

	Cu	Ni	Mn	Zn	Pb	Cd	Cr	Fe	pН	Conductivity	Total hardness
Ni	-0.041	1									
Mn	0.820*	-0.195	1								
Zn	0.823*	0.190	0.639	1							
Pb	0.674	-0.410	0.808*	0.382	1						
Cd	0.643	-0.170	0.729*	0.726*	0.525	1					
Cr	0.837*	-0.181	0.985*	0.643	0.798*	0.626	1				
Fe	0.446	-0.387	0.135	0.428	0.242	0.196	0.177	1			
pН	-0.122	-0.540	0.394	-0.193	0.259	0.170	0.375	-0.099	1		
Conductivity	0.334	-0.313	0.420	0.567	0.342	0.471	0.433	0.646	0.417	1	
Total hardness	-0.831*	0.228	-0.761*	-0.626	-0.528	-0.389	-0.834	-0.322	-0.157	-0.306	1

 Table 4. Pearson's correlation matrix for the physicochemical parameters and HMs content in water from R.

 Manafwa during the wet season.

* Significant at the 0.05 level (2-tailed).

For the dry season, Pearson's bivariate correlation coefficients (Table 5) indicated a positive correlation between Cu and Mn (p = 0.03), Fe and pH (p = 0.10), and significant negative correlations between Cu and total hardness (p = 0.038), and Mn and total hardness (p = 0.027). The positive correlation indicates that Cu and Mn entered the river through the same anthropogenic routes, and their concentrations would increase or decrease proportionately [16, 72]. Fe is associated with pH because its solubility in aqueous matrices is pH-dependent [73]. These results were also in agreement with PCA results (Figure 12). Overall, multivariate statistical analyses indicated that the comparatively higher pollution levels experienced in R. Manafwa during the wet season are due to anthropogenic contributions.



Figure 11. Principal Component Analysis plot showing the effect of three components influencing the variation of HMs in water from R. Manafwa during the wet season.

 Table 5. Pearson's correlation matrix for the physicochemical parameters and HMs content in water from R.

 Manafwa during the dry season.

					0	,			
	Cu	Ni	Mn	Zn	Pb	Fe	pН	Conductivity	Total hardness
Cu	1								
Ni	-0.129	1							
Mn	0.891^{*}	-0.062	1						
Zn	0.390	0.601	0.252	1					
Pb	-0.025	-0.232	0.001	-0.114	1				
Fe	-0.431	0.080	-0.061	-0.490	0.269	1			



Figure 12. Principal Component Analysis plot showing the effect of three components influencing the variation of HMs in water from R. Manafwa during the dry season.

3.4. Health risks assessment results.

In the wet season, the average daily doses through ingestion of contaminated water ranged from 0.120×10^{-3} mg/kg/day (for Cd ingested by children at A₁, A₂, B₁, and B₂) to 138.95×10^{-3} mg/kg/day for Fe ingested in water from site M4 by adults (Table S1). The daily doses through dermal contact ranged from 0.0010×10^{-8} mg/kg/day for adults exposed to Cd at B₂ to 258.2318×10^{-8} mg/kg/day for adults exposed to Fe at M₄. In the dry season, the average daily doses through ingestion of contaminated water span from 0.006×10^{-3} mg/kg/day (for Zn ingested by children at A₁, A₂, and B₁) to 154.200×10^{-3} mg/kg/day (for Cu ingested in water from site M4 by children). On the other hand, the estimated daily doses through dermal contact varied between 0.0037×10^{-8} mg/kg/day (for Cr adsorbed by the skin of children at M₄) to 399.99×10^{-8} mg/kg/day for Fe adsorbed onto adult's skin at A₁. The THO through ingestion of the HMs, and hence the total THQ for the foregoing HMs exceeded 1 during the wet season at B₂ and M₁ for Ni ingested by children and at all sites for Ni ingested by adults. Similarly, the THQ for ingesting the HMs during the dry season, and hence the total THQ for the HMs exceeded 1 for Ni and Cu ingested by children at A₂, M₁, and M₄, respectively (Table S1). All the THQ from dermal contact were less than 1, and varied between 0.0005×10^{-6} (for Fe at B₁ adsorbed by the skin of children) and 137.3067×10^{-6} (for Zn at M₄ adsorbed by the skin of adults) for the wet season, and 0.0003×10^{-6} (for Mn at A₁ adsorbed by the skin of children) to 21.3067×10^{-6} (for Zn at M₁ adsorbed by the skin of adults) for the dry season. Since some of the average daily doses for Ni and Cu were higher than the corresponding reference doses for ingestion of contaminated water (THQ >1), non-carcinogenic health risks may result from water consumption from the sampled stations of R. Manafwa.

For carcinogenic health risks, the cancer risk values ranged from 4.4625 $\times 10^{-10}$ to 2.22075 $\times 10^{-6}$, with lower cancer risk values in the dry season than wet season (Table S2). For carcinogenic health risks, the borderline given by US EPA is 1×10^{-4} to 1×10^{-6} and is considered unacceptable where the risk surpasses 1×10^{-4} . The cancer risk values (and the total cancer risk values) obtained for the HMs per sampling site in this study did not surpass 1×10^{-4} , suggesting that no potential cancer risks could arise from water consumption from R. Manafwa, Uganda.

Overall, the health risks assessment performed in this study was based on the HMs analyzed. It is recommended that risk characterizations should be cumulative to account for aggregate exposures to multiple compounds or mixtures causing similar toxicological effects [74]. This is achieved using Adversity Specific Hazard Index for Cumulative risk assessment suitable for toxicants with multiple residues that exert similar toxicological effects, such as pesticides and polychlorinated biphenyls.

4. Conclusions

This study showed that R. Manafwa in Eastern Uganda is contaminated with HMs at concentrations well below the maximum WHO guidelines for drinking water. However, there are discernable non-carcinogenic health risks that may arise from ingesting water from the river. Children are at higher cancer risks than adults, and in all cases, the contribution of Ni (in both seasons) and Cu (during the dry season) towards the target hazard quotient was significant. The heavy metal Cr is the main driver of potential carcinogenicity in the incremental lifetime cancer risk assessment results for the wet season, while Pb is the sole driver in the dry season. Thus, regulatory authorities should intervene and reduce the pollution of R. Manafwa by strengthening restrictions on sand mining and the dumping of waste into the river. Further studies should examine the spatial variations in the concentrations of the HMs in sediments and biota.

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Conflicts of Interest

The authors declare no conflict of interest.

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Supplementary materials

 Table S1. Estimated daily doses and hazard quotients through dermal contact and ingestion of water from River

 Manafwa, Uganda.

Season	Group	Site	Metal	$\begin{array}{c} \text{ADD}_{\text{Ingestion}} \\ (\times 10^{-3} \\ mg/kg/day) \end{array}$	THQ Ingestion	ADD Dermal contact (× 10 ⁻⁸ mg/kg/day)	THQ Dermal contact ($\times 10^{-6}$)
			Cu	<u>3 480</u>	0.0870	1.0827	1.0827
			Ni	0.180	0.600	0.0560	1.0370
			Mn	1.320	0.0044	0.4107	0.0042
			Zn	0.960	0.0320	0.0299	0.0498
		A_1	Pb	0.780	0.0195	0.0243	0.4629
			Cd	0.120	0.1200	0.0037	0.6167
			Cr	0.480	0.3200	0.00149	0.2483
			Fe	30.120	0.0430	9.37070	0.0067
			Cu	2.760	0.0690	0.85870	0.42935
			Ni	0.540	0.1800	0.0168	0.03111
			Mn	1.200	0.0040	0.3733	0.00389
		A ₂	Zn	1.800	0.0600	0.5600	9.33334
		112	Pb	0.336	0.0840	0.01045	0.19905
			Cd	0.180	0.1800	0.05600	9.33334
			Cr	0.360	0.00025	0.01120	1.86667
			Fe	26.760	0.0383	8.3253	0.00059
			Cu	4.080	0.1020	0.12693	0.12693
			N1	0.180	0.0600	0.05970	0.11056
			Mn Zn	1.320	0.0044	0.4107	0.00428
		B_1	Zn	1.440	0.0480	0.4480	8.12334
			PD Cd	0.420	0.1050	0.0131	6 8500
			Cu	0.120	0.1200	0.0411	1.8667
			CI Ee	0.300	0.00024	0.0112	0.0005
				4 320	0.0330	2 4102	2 4102
			Ni	4.320	3 0000	2.4192	2.4192
			Mn	1 560	0.0052	0.8736	0.0091
			Zn	1.500	0.0032	0.8750	13 440
	Children	B_2	Ph	0.360	0.0900	0.02016	0 3840
			Cd	0.120	0.1200	0.02010	1 1 2 0 0
			Cr	0.480	0.00032	0.00269	0.4483
Wet season			Fe	32.640	0.04663	18.2784	0.0013
			Cu	3.960	0.0990	2.2176	2.2176
			Ni	0.600	2.0000	0.3360	0.6223
			Mn	2.160	0.0072	1.2096	0.0126
		м	Zn	1.320	0.0440	0.7392	12.320
		1011	Pb	0.780	0.1950	0.04368	0.8320
			Cd	0.180	0.1800	0.01008	1.6800
			Cr	0.360	0.00024	0.02016	3.3600
			Fe	26.520	0.03789	14.8512	0.0011
			Cu	5.160	0.1290	0.28896	0.28896
			Ni	1.140	0.3800	0.63840	1.1823
			Mn	2.400	0.0080	1.3440	0.0140
		M2	Zn	3.240	0.1080	1.8144	30.240
		2	Pb	0.480	0.1200	0.02688	0.5120
			Cd	0.180	0.1800	0.01008	1.6800
			Cr	0.480	0.0003	0.02688	4.4800
			Fe	27.00	0.0386	15.1200	0.0011
			Cu N:	5.400	0.1350	1.6800	1.6800
			IN1 Mn	0.180	0.0600	0.0560	0.1057
			Zn	2.280	0.0700	0.7093	0.0074
		M_3	ZII Ph	2.700	0.0920	0.0307	0 3005
			Cd	0.000	0.1800	0.0205	0.9334
			Cr	0.180	0.0003	0.0149	2 4834
			Fe	109 08	0.1558	33 936	0.0024
			Cu	7.200	0.1800	2.2400	2.2400
			Ni	0.300	1.0000	0.0933	0.1728
			Mn	0.164	0.0006	6.1227	0.0638
		M_4	Zn	19.680	0.6560	1.0827	18.0450
			Pb	1.200	0.3000	0.0373	0.7105
			Cd	0.240	0.2400	0.0075	1.2500

a a		<u>.</u>		ADD Ingestion		ADD Downel contact	ТНО
Season	Group	Site	Metal	$(\times 10^{-3})$	THQ Ingestion	$(\times 10^{-8} \text{ mg/kg/day})$	Dermal contact (× 10 ⁻⁶
			Cr	<u>mg/kg/day)</u> 1.320	0.0009	0.0411	6.8500
			Fe	47.640	0.0681	14.821	0.0011
_			Cu	10.150	0.2538	8.2380	8.2380
			Ni	0.525	1.7500	0.4261	0.0079
			Mn	3.850	0.1283	3.1249	0.0326
		A_1	Zn	2.800	0.0933	2.2727	37.8784
			Pb	2.275	0.5688	0.1847	3.5181
			Ca Cr	0.350	0.3500	0.0284	4./334
			Ee Ee	1.400 87 850	0.1255	71 305	0.0051
				8 050	0.1255	0.6534	0.6534
			Ni	1.575	5.2500	1 2784	2 3674
			Mn	3.500	0.0117	2.8408	0.0295
			Zn	5.250	0.1750	4.2613	71.0217
		A_2	Pb	0.980	0.2450	0.0795	1.5143
			Cd	0.525	0.5250	0.0426	7.1000
			Cr	0.105	0.0700	0.08523	14.2050
		_	Fe	81.550	0.1165	63.3506	0.00453
			Cu	11.900	0.2975	9.6588	9.6588
			Ni	0.525	1.7500	0.42613	0.7891
			Mn	3.850	0.0128	3.1249	0.0326
		Bı	Zn	4.200	0.1400	3.4090	56.8167
		-1	Pb	1.225	0.3063	0.0994	1.8934
			Cd	0.350	0.3500	0.0284	4.7334
			Cr E-	1.050	0.0007	0.0852	14.200
			Fe	12 600	0.0980	10 227	10.0397
	Adults		Cu Ni	2 625	0.5150 8 7500	2 1306	3 9456
	Adults		Mn	4 550	0.0152	3 6931	0.0385
			Zn	4 200	0.1400	3 4090	56 8167
		B_2	Pb	1.050	0.2625	0.0853	1.6248
			Cd	0.350	0.3500	0.0010	0.1667
			Cr	1.400	0.00093	0.0284	4.7334
			Fe	95.200	0.13600	77.271	0.0055
			Cu	11.550	0.28875	9.3748	9.3748
			Ni	1.750	5.83333	1.4204	2.6304
			Mn	2.160	0.0072	5.1135	0.0533
		M_1	Zn	6.300	0.2100	3.1249	52.0817
			Pb	2.275	0.56875	0.1847	3.5181
			Cd	0.525	0.52500	0.0426	7.1000
			Cr Ea	1.050	0.00070	0.0852	14.200
			Cu	15.050	0.27625	12 216	12 2160
			Cu Ni	3 325	0.57025 11 0833	2 6988	12.2100
			Mn	5.525 7 00	0.02333	2.0900	4.2270
		_	Zn	3.240	0.10800	7.6703	127.8334
		M_2	Pb	1.400	0.35000	0.1136	2.1638
			Cd	0.525	0.52500	0.0426	7.1000
			Cr	1.400	0.00093	0.1136	18.9334
			Fe	78.750	0.11250	63.919	0.0046
		_	Cu	15.750	0.39375	12.784	12.784
			Ni	0.525	1.31250	0.4261	7.8907
			Mn	6.650	0.02117	5.3975	0.0562
		M 3	Zn	8.050	0.26833	6.5339	108.8984
			Pb	1.925	0.48125	0.1562	2.9750
			Cd Cr	0.525	0.52500	0.0426	/.1000
				1.400	0.00093	0.1130	18.9334
		М		21 00	0.43300	230.2318	17 0/50
		1 v1 4	Ni	21.00 0 875	0.52500 2 01667	0.7102	17.0450
			Mn	57 400	0 19133	465 897	1.31 <i>32</i> 4 8531
			Zn	10 150	0 38333	8 2384	137 3067
			Ph	3.500	0.87500	0.2841	5,4114
			Cd	0.700	0.7000	0.0568	9.4667
			Cr	3.850	0.00233	0.3125	52.0834
			Cr Fe	3.850 138.95	0.00233 0.19850	0.3125 112.7811	52.0834 0.0081

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Season	Group	Site	Metal	ADD Ingestion ($\times 10^{-3}$ mg/kg/day)	THQ Ingestion	ADD Dermal contact (× 10 ⁻⁸ mg/kg/day)	THQ Dermal contact (× 10 ⁻⁶)
			Ni	0.1200	0.4000	0.0373	0.0691
			Mn	8 6400	0.0288	0.0269	0.0003
			Zn	0.1800	0.0060	0.0560	0.9334
			Pb	1.020	0.2550	0.0317	0.6038
			Cd	NA	NA	NA	NA
			Cr	NA	NA	NA	NA
			Fe	168.96	0.2414	52.565	0.0038
			Cu	3.500	0.0875	0.1307	0.1307
			Ni	0.420	1.4000	0.0373	0.0691
			Mn	13.200	0.0440	4.1067	0.0428
		٨	Zn	0.1800	0.0060	0.0560	0.9334
		A_2	Pb	0.6000	0.1500	0.0187	0.3562
			Cd	NA	NA	NA	NA
			Cr	NA	NA	NA	NA
			Fe	140.76	0.20109	43.792	0.0031
			Cu	0.36	0.0090	0.0120	0.0120
			Ni	0.18	0.6000	0.0560	0.1037
			Mn	17.28	0.0576	5.3760	0.0560
		Bı	Zn	0.18	0.0060	0.0560	0.9334
		DI	Pb	0.72	0.1800	0.0224	0.4267
			Cd	NA	NA	NA	NA
			Cr E-	NA	NA 0.2412	NA 52.527	NA 0.00275
			Fe	168.84	0.2412	52.527	0.00375
				0.360	0.0090	0.0112	0.0112
			IN1 Mm	0.120	0.4000	0.03/3	0.0091
			Zn	0.500	0.0212	1.9787	0.0206
		\mathbf{B}_2	ZII Dh	0.180	0.0000	0.0300	0.9334
			Cd	0.000 N A	0.1500 N A	0.0187 NA	0.3302 NA
			Cr	NA	NA	NA	NA
			Fe	153.24	0 2189	47 675	0.0034
			Cu	0.4200	0.0105	0.1307	0.1307
			Ni	0.6000	2.0000	0.1867	0.3457
			Mn	16.200	0.0540	5.0400	0.0525
		м	Zn	0.5400	0.0180	0.1680	2.800
		\mathbf{M}_1	Pb	0.6600	0.1650	0.0205	0.3905
			Cd	NA	NA	NA	NA
			Cr	NA	NA	NA	NA
			Fe	151.680	0.2167	47.1893	0.0034
			Cu	0.4800	0.0120	0.1493	0.1493
			Ni	0.1200	0.4000	0.0373	0.0691
			Mn	16.680	0.0556	5.1893	0.0541
		M_2	Zn	0.1800	0.0060	0.0560	0.9334
			Pb	0.7800	0.1950	0.0243	0.4629
			Cd Cr				NA
			Cr Ea	INA 151 44	INA 0.2162	INA 47.1147	INA 0.0034
			Cu	0 5/00	0.2105	<u>+/.114/</u> 0.1680	0.0034
			Ni	0.3400	0.0135	0.1000	0.1000
			Mn	30 720	0.4000	9 5570	0.0091
			Zn	0 4200	0.0140	0 1 3 0 7	2 1784
		M_3	Pb	0.7200	0.1800	0.0224	0.4267
			Cd	NA	NA	NA	NA
			Cr	NA	NA	NA	NA
			Fe	1.285	0.00184	47.9733	0.0034
			Cu	154.200	3.8550	0.1680	0.1680
			Ni	0.1200	0.4000	0.0373	0.0691
			Mn	31.440	0.1408	9.7813	0.1019
		M4	Zn	0.4200	0.0140	0.1307	2.1700
		1717	Pb	0.7200	0.1800	0.0224	0.4267
			Cd	NA	NA	NA	NA
			Cr	0.1200	0.00008	0.0037	0.6167
			Fe	154.20	0.2203	47.973	0.0034
			Cu	0.1050	0.00263	0.8523	0.8523
	Adults	A_1	IN1 Mrs	0.0350	0.1106/	0.2841	0.5201
			7n	0.0525	0.00840	20.434 0.4261	7 1017
			Z 11	0.0525	0.001/5	0.4201	/.101/

				ADD Ingestion		4.0.0	
Season	Group	Site	Metal	(× 10 ⁻³	THQ Ingestion	ADD Dermal contact	THQ
	•			mg/kg/dav)		$(\times 10^{\circ} \text{mg/kg/day})$	Dermal contact (× 10 °)
			Pb	0.2975	0.07438	0.2415	4.6000
			Cd	NA	NA	NA	NA
			Cr	NA	NA	NA	NA
			Fe	49.280	0.0704	399.99	0.0285
			Cu	1.2250	0.0306	0.9943	0.9943
			Ni	0.0350	0 1 1 6 6 7	0 2841	0.5261
			Mn	3.8500	0.01283	31.249	0.3255
			Zn	0.0525	0.00175	0 4261	7 1017
		A_2	Ph	0.1750	0.00175	0.14201	2 7048
			Cd	NA	NA	NA	NA
			Cr	NΔ	NΔ	NΔ	NΔ
			Fe	41.055	0.05865	3 3323	0.00023
				0 1050	0.00000	0.8523	0.8523
			Ni	0.1050	0.1750	0.4261	0.7891
			Mn	176.40	0.5880	40.908	0.7671
			Zn	1 8375	0.06125	40.000	7 1017
		B_1	Dh	0.2100	0.00125	0.4201	3 2476
			Cd	0.2100 NA	0.05000 NA	0.1705 NA	5.2470 NA
			Cu	NA NA	NA	NA	NA
			Ea	10 245	0.07025	200 705	0.02855
			Cu	49.243	0.07033	0.8522	0.02833
				0.1030	0.002023	0.0323	0.6525
			INI Me	0.0550	0.11007	0.2041	0.3201
			Zn	1.8330	0.00018	13.030	0.1308
		B_2	ZII Dh	0.0323	0.00175	0.4201	2 7055
			PD	0.1750	0.04373 NA	0.14204 NA	2.7055 NA
			Cu Cr	INA NA	INA NA	INA NA	INA
			Cr Ea	NA 44.605	NA 0.06424	NA 262 777	NA 0.2501
			re Cu	44.095	0.00424	0.0042	0.2391
				0.1223	0.00500	0.9945	0.9945
			INI Mm	0.1730	0.36333	1.4204	2.0304
			7	4.7230	0.01373	38.330	0.5995
		M_1	ZII	0.1373	0.00323	1.2784	21.5007
			PD	0.1925	0.04615	0.1302	2.8920
			Ca Ca	NA NA	INA NA	NA NA	NA
			Cr E-	INA 44.240	NA 0.0(220	NA 250.091	NA 0.0257
			Fe	44.240	0.06320	559.081	0.0257
			Cu	0.1400	0.00350	1.1363	1.1363
			IN1 M	0.0350	0.1166/	0.2841	0.5261
			ivin 7	4.8050	0.01622	39.488	0.4114
		M_2	Zn D'	0.0015	0.00005	0.4261	/.101/
			PD	0.0525	0.01313	0.184/	3.5181
			Cđ	NA	NA	NA	NA
			Cr	NA 44.170	NA 0.06210	NA 259 512	NA 0.0257
			ге	44.170	0.00310	338.313	0.0250
			Cu	0.15/5	0.00394	1.2/84	1.2/84
			N1	0.0350	0.11667	0.2841	0.52611
			Mn	8.9600	0.029867	12.125	0.75755
		M ₃	Zn	0.1225	0.004083	0.9943	16.5/17
		2	Pb	0.2100	0.05250	0.1705	3.24762
			Cd	NA	NA	NA	NA
			Cr	NA	NA	NA	NA
			Fe	44.975	0.06425	365.047	0.00267
			Cu	0.1575	0.00394	1.278	1.27800
			Ni	0.0350	0.11667	0.2841	0.52611
			Mn	9.1700	0.03057	74.430	0.77531
		M_4	Zn	0.1225	0.00408	0.9943	16.5717
		1.14	Pb	0.2100	0.05250	0.1705	3.24762
			Cd	NA	NA	NA	NA
			Cr	0.0350	0.000023	0.0284	4.73334
			Fe	44.975	0.064250	365.047	0.02607

Values in **bold** indicate exceedance of reference dose for estimated daily doses (ADD Ingestion or ADD Dermal contact) or unity for target hazard quotient (THQ). NA = Not applicable. The oral reference doses are 4.0×10^{-2} , 3.0×10^{-4} , 3.0×10^{-2} , 3.0×10^{-2} , 4.0×10^{-3} , 1.0×10^{-3} , 1.5×10^{0} and 7×10^{-1} mg/kg/day for Cu, Ni, Mn, Zn, Pb, Cd, Cr and Fe, respectively. The corresponding reference doses for Cu, Ni, Mn, Zn, Pb, Cd, Cr and Fe are 1.0×10^{-2} , 5.40×10^{-3} , 9.6×10^{-1} , 6.0×10^{-4} , 6.0×10^{-5} , 6.0×10^{-5} , and 1.4×10^{2} mg/kg/day, respectively.

		s unoug			narwa, Uganua.	
Season	Group	Site	Metal	Cancer risk	Total Cancer Ris	
		٨	Pb Cd	0.03×10^{-7}	2 8676 -10-6	
		\mathbf{A}_1	Cu	4.30×10^{-7}	2.8020 ×10	
			Dh	2.4 ×10		
		4.2	FU Cd	2.830×10^{-8}	8 8536 ×10 ⁻⁸	
		\mathbf{A}_{2}	Cr	1.728×10^{-8}	0.0550 ×10	
			Ph	3 57 ×10-9		
		B.	Cd	5.57×10^{-8}	6 645 ×10 ⁻⁸	
		\mathbf{D}_1	Cu	4.30×10^{-8}	0.045 ×10	
			Dh	3.06 ×10 ⁻⁹		
		D.		5.00×10	7 166 × 10-7	
		D 2	Cu	4.50×10^{-9}	7.100×10	
	Children		Dh	2.304×10°		
		М.	FU	6.03×10^{-8}	$2,40062\times 10^{-7}$	
		1111	Cu	0.84×10^{-7}	2.49003 ×10	
				1.80 ×10		
		м	PD	4.08×10^{-9}	2 00000 10-7	
		11/12	Ca	6.84×10^{-3}	3.08808 ×10 ⁷	
			Cr	2.40 ×10 ⁷		
			PD	5.61×10^{-8}	a and <i>i</i> a a 7	
		M 3	Ca	$0.84 \times 10^{\circ}$	3.08961 ×10-7	
			Cr	2.40×10-7		
			Pb	5.61 ×10-9		
		M_4	Cd	9.12 ×10 ⁻⁸	7.51761 ×10-7	
Wet season			Cr	6.60 ×10-7		
			Pb	1.934 ×10-8		
		A_1	Cd	1.33 ×10-7	8.5234 ×10-7	
		-	Cr	7.0 ×10-7		
			Pb	8.33 ×10-9		
		A_2	Cd	1.995 ×10-7	2.8283 ×10-7	
			Cr	7.50 ×10-8		
			Pb	1.041×10^{-8}		
		B_1	Cd	1.33×10^{-7}	6.6841 ×10-7	
			Cr	5.25 ×10-7		
		_	Pb	8.925 ×10-9	7	
		B_2	Cd	1.33×10-7	8.41925 ×10-7	
	Adults		Cr	7.00 ×10-7		
	7 iduito		Pb	1.934 ×10 ⁻⁸	-	
		M_1	Cd	1.995×10-7	7.4384×10^{-7}	
			Cr	5.250 ×10-7		
			Pb	1.19 ×10 ⁻⁹	_	
		M_2	Cd	1.995×10 ⁻⁷	9.0069 ×10-7	
			Cr	7.00 ×10 ⁻⁷		
		M ₃	Pb	1.636 ×10 ⁻⁸		
			Cd	1.995×10 ⁻⁷	9.1586×10^{-7}	
			Cr	7.00 ×10 ⁻⁷		
		M_4	Pb	2.975 ×10 ⁻⁸		
			Cd	2.66 ×10 ⁻⁷	2.22075 ×10 ⁻⁶	
			Cr	1.925 ×10-6		
			Pb	8.67 ×10 ⁻⁹		
		A_1	Cd	NA	8.67 ×10 ⁻⁹	
			Cr	NA		
			Pb	5.10 ×10 ⁻⁹		
		A_2	Cd	NA	5.10 ×10 ⁻⁹	
		-	Cr	NA		
			Рh	6 12 v10-9		
		P.	ru	0.12 ×10	6 12 v10-9	
Dry saason	Children	DI	Cd	NA	0.12 ×10	
Dry season	Cinidren		Cr	NA		
			Pb	5.10×10 ⁻⁹		
		B_2	Cd	NA	5.10 ×10 ⁻⁹	
			Cr	NA		
			Pb	5.61 ×10 ⁻⁹		
		M_1	Cd	NA	5.61 ×10 ⁻⁹	
			Cr	NA		
		14	Pb	6.63 ×10 ⁻⁹	C (2 10. ⁰	
		M_2	Cd	NA	6.63 ×10-2	
		-				

 Table S2. Cancer risks through ingestion of water from River Manafwa, Uganda.

Season	Group	Site	Metal	Cancer risk	Total Cancer Ris
			Cr	NA	
			Pb	6.12 ×10 ⁻⁹	
		M_3	Cd	NA	6.12 ×10 ⁻⁹
			Cr	NA	
			Pb	6.12 ×10 ⁻⁹	
		M4	Cd	NA	6.612 ×10 ⁻⁸
			Cr	6.00 ×10 ⁻⁸	
			Pb	2.52875 ×10-9	
		A_1	Cd	NA	2.52875 ×10-9
			Cr	NA	
			Pb	1.4875 ×10-9	
		A_2	Cd	NA	1.4875 ×10 ⁻⁹
			Cr	NA	
			Pb	1.785 ×10 ⁻⁹	
		\mathbf{B}_1	Cd	NA	1.785 ×10 ⁻⁹
			Cr	NA	
			Pb	1.4875 ×10 ⁻⁹	
	Adults	\mathbf{B}_2	Cd	NA	1.4875 ×10 ⁻⁹
			Cr	NA	
			Pb	1.636 ×10 ⁻⁹	
		M_1	Cd	NA	1.636 ×10-9
			Cr	NA	
			Pb	4.4625 ×10 ⁻¹⁰	
		M_2	Cd	NA	4.4625 ×10 ⁻¹⁰
			Cr	NA	
			Pb	1.785 ×10 ⁻⁹	
		M_3	Cd	NA	1.785 ×10-9
			Cr	NA	
		-	Pb	1.785 ×10 ⁻⁹	
		M_4	Cd	NA	1.9285 ×10 ⁻⁸
			Cr	1.750 ×10 ⁻⁸	

The ingestion cancer slope factors (mg/kg/day) for Pb, Cr and Cd are 8.5×10^{-6} , 5.0×10^{-4} and 3.8×10^{-4} , respectively.